

# Short-term Effects of Ambient Gaseous Pollutants and Particulate Matter on Daily Mortality in Shanghai, China

Guohai Chen<sup>1</sup>, Guixiang Song<sup>2</sup>, Lili Jiang<sup>2</sup>, Yunhui Zhang<sup>3</sup>, Naiqing Zhao<sup>4</sup>, Bingheng CHEN<sup>3</sup> and Haidong KAN<sup>3</sup>

<sup>1</sup>Shanghai Environmental Monitoring Center, <sup>2</sup>Shanghai Municipal Center of Disease Control and Prevention, <sup>3</sup>Department of Environmental Health and <sup>4</sup>Department of Health Statistics, School of Public Health, Fudan University, China

**Abstract: Short-term Effects of Ambient Gaseous Pollutants and Particulate Matter on Daily Mortality** in Shanghai, China: Guohai CHEN, et al. Shanghai Environmental Monitoring Center, China-Identification of the specific pollutants contributing most to the health hazard of the air pollution mixture may have important implications for environmental and social policies. In the current study, we conducted a time-series analysis to examine the specific effects of major air pollutants [particulate matter less than 10 microns in diameter (PM<sub>10</sub>), sulfur dioxide (SO<sub>2</sub>), and nitrogen dioxides (NO<sub>2</sub>)] on daily mortality in Shanghai, China, using both single-pollutant and multiple-pollutant models. In the single-pollutant models, PM<sub>10</sub>, SO<sub>2</sub>, and NO, were found to be associated with mortality from both all non-accidental causes and from cardiopulmonary diseases. Unlike some prior studies in North America, we found a significant effect of gaseous pollutants (SO<sub>2</sub> and NO<sub>2</sub>) on daily mortality even after adjustment for PM<sub>10</sub> in the multiple-pollutant models. Our findings, combined with previous Chinese studies showing a consistent, significant effect of gaseous pollutants on mortality, suggest that the role of outdoor exposure to SO2 and NO2 should be investigated further in China.

(J Occup Health 2008; 50: 41–47)

Key words: Outdoor air pollution, Gaseous pollutants, Particle, Time-series

linked to adverse health effects, including increased mortality, increased rates of hospital admissions and emergency department visits, exacerbation of chronic

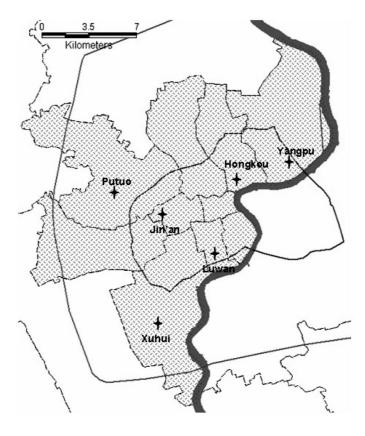
Short term exposure to outdoor air pollution has been

respiratory conditions (e.g., asthma), and decreased lung function<sup>1)</sup>. Recent multi-city analyses conducted in the U.S., Canada and Europe provide further evidence supporting the coherence and plausibility of these associations<sup>2-6)</sup>. Most of these studies were conducted in developed countries, and only a small number of studies have been conducted in Asia7). The need remains for studies of cities in developing countries, where characteristics of outdoor air pollution (e.g. air pollution level, chemical composition and size of particles, and fate and transport of pollutants), meteorological conditions and socio-demographic patterns may differ from North America and Western Europe.

Ambient air pollution is a complex mixture composed of both solid particles and gaseous pollutants. Although the strongest evidence linking outdoor air pollutants with adverse health effects is for solid particulates (e.g. PM<sub>10</sub>), many researchers have reported associations for gaseous pollutants<sup>5,8)</sup>. In China, for example, one study conducted by Xu et al. showed that it was SO<sub>2</sub>, not total suspended particle (TSP), that was associated with mortality increase in Beijing<sup>9)</sup>. Similar results have also been presented by the studies conducted worldwide, which have led some authors to conclude that the pollutants measured and included in models of daily mortality might be better interpreted as indicators of the biologically relevant pollutant mixture and that the best indicators might vary in different geographic areas<sup>10)</sup>. Obviously, identification of the specific pollutants contributing most to the health hazard of the air pollution mixture may have important implications for environmental and social policies, and for local government in taking steps to protect the total population in general, and the sensitive population in

In the current study, we conducted a time-series analysis to examine the effects of major air pollutants [particulate matter less than 10 microns in diameter

School of Public Health, Fudan University, Shanghai 200032, China (e-mail address: haidongkan@gmail.com)



**Fig. 1.** Map of research area and location of six monitoring stations in Metropolitan Shanghai.

 $(PM_{10})$ , sulfur dioxide  $(SO_2)$ , and nitrogen dioxides  $(NO_2)$ ] on daily mortality in Shanghai, China, using both single-pollutant and multiple-pollutant models.  $PM_{10}$ ,  $SO_2$  and  $NO_2$  are criteria pollutants in China and have been regularly monitored since 1996.

# Methods

Shanghai, the most populous city in China, is located at the tip of the Yangtze River Delta in eastern China. The city comprises urban/suburban districts and counties, with a total area of 6,341 square kilometres (km²), and had a population of 13.1 million at the end of 2004, representing about 1% of China's total. Our study area was limited to the traditional nine urban Districts of Shanghai (289 km²)—Huangpu, Jinan, Luwan, Xuhui, Yangpu, Changnin, Yangpu, Putuo and Zhabei. The target population included all permanent residents living in the area, around 6.3 million in 2004.

Daily mortality data (excluding accidents & injuries) of residents living in the nine urban districts of Shanghai from Jan. 1, 2001 to Dec. 31, 2004 were collected from the database of Shanghai Municipal Center of Disease Control and Prevention (SMCDCP). The death report system in Shanghai was implemented in 1951, and has

been computerized since 1990. For both at-home and hospital deaths, physicians complete death certificate cards. The information on the cards is then sent to SMCDCP through a computer network. The database for 2001 and 2002–2004 was coded according to the International Classification of Diseases, Revision 9 (ICD-9) and 10 (ICD 10), respectively. The mortality data were classified into deaths due to all non-accidental causes (ICD-9 <800; ICD-10 A00-R99), cardiovascular diseases (ICD-9 390-459; ICD-10 I00-I99), and respiratory diseases (ICD-9 460-519; ICD-10 J00-J98).

Daily data on levels of PM<sub>10</sub>, SO<sub>2</sub> and NO<sub>2</sub> from Jan. 1, 2001 to Dec. 31, 2004 in Metropolitan Shanghai were retrieved from the database of the Shanghai Environmental Monitoring Center (SEMC). The daily concentrations for each pollutant were averaged from the available monitoring results of six fixed-site stations located in urban areas of Shanghai and covered by China National Quality Control (Fig. 1). The stations are mandated not to be in direct proximity to traffic, industrial sources, buildings or residential sources of emissions from the burning of coal, waste, or oil. Thus, our monitoring results reflect the background air pollution levels in Shanghai.

|                               | Mean ± SD        | Min  | 25th percentile | Median | 75th percentile | Max   |
|-------------------------------|------------------|------|-----------------|--------|-----------------|-------|
| Daily death counts            |                  |      |                 |        |                 |       |
| Total (non-accident)          | $119.0 \pm 22.5$ | 51.0 | 103.0           | 115.0  | 133.0           | 198.0 |
| Cardiovascular                | $44.2 \pm 11.0$  | 11.0 | 36.0            | 43.0   | 51.0            | 85.0  |
| Respiratory                   | $14.3 \pm 6.4$   | 3.0  | 10.0            | 13.0   | 17.0            | 45.0  |
| Air pollutants concentrations |                  |      |                 |        |                 |       |
| $PM_{10} (\mu g/m^3)$         | $102.0 \pm 64.8$ | 14.0 | 56.3            | 84.0   | 128.3           | 566.8 |
| $SO_2 (\mu g/m^3)$            | $44.7 \pm 24.2$  | 8.4  | 27.5            | 40.0   | 56.2            | 183.3 |
| $NO_2 (\mu g/m^3)$            | $66.6 \pm 24.9$  | 13.6 | 50.2            | 62.5   | 79.2            | 253.7 |
| Meterologic measures          |                  |      |                 |        |                 |       |
| Mean Temperature (°C)         | $17.7 \pm 8.5$   | -2.4 | 10.3            | 18.3   | 24.7            | 34.0  |
| Relative humidity (%)         | $72.9 \pm 11.4$  | 33.3 | 65.5            | 73.5   | 81.0            | 97.0  |

Table 1. Summary statistics of daily death numbers, air pollutant concentrations and weather conditions in Shanghai (2001–2004)

To allow adjustment for the effect of weather conditions on mortality, meteorological data (daily minimum, maximum and mean temperature, relative humidity, and dew point temperature) from Jan. 1, 2001 to Dec. 31, 2004 were obtained from the database of the Shanghai Meteorological Bureau (SMB).

All daily mortality, pollutant and meteorological data were validated by an independent auditing team assigned by the Health Effects Institute (HEI), the funding agency of this study.

### Statistical analysis

The daily death, air pollution and weather data are linked by date and therefore can be analyzed with a time-series design<sup>11</sup>. Because counts of daily mortality data typically follow a Poisson distribution and the relations between mortality and explanatory variables are mostly non-linear<sup>11</sup>, the core analysis was a generalized additive model (GAM) with log link and a Poisson error that accounted for smooth fluctuations in daily mortality.

We first built basic models for various mortality outcomes not including the air pollution and weather variables. We incorporated smoothed spline functions of time, which can accommodate non-linear and nonmonotonic patterns between mortality and time, offering a flexible modeling tool<sup>12</sup>). Day of the week (DOW) was also included in the basic models. In our analysis, a partial autocorrelation function (PACF) was used to guide the selection of degrees of freedom (df)13). Specifically, we used 4-6 df per year for time trend. When the absolute magnitude of the PACF plot was less than 0.1 for the first two lag days, the basic model was regarded as adequate; if this criteria was not met, auto-regression (AR) terms for lag up to 7 days was introduced to improve the model. In this way, 4, 4 and 5 df per year for time trend, as well as 3, 2 and 4 lag-day AR terms, were used in our basic models for total, cardiovascular and respiratory

mortality, respectively.

After we established the basic models, we introduced the pollutant and weather variables and analyzed their effects on mortality outcomes. Based on the previous literature<sup>6, 14, 15)</sup>, 3 df (whole period of study) for temperature and relative humidity could control well for their effects on mortality and was used in the model.

Briefly, we fitted the following log-linear GAM to obtain the estimated pollution log-relative rate  $\beta$  in Shanghai:

$$logE(Y_t) = \beta Z_t + DOW + ns(time,df) + ns(temperature,3) + ns(humidity,3) + intercept$$

Here  $E(Y_t)$  represents the expected number of deaths on day t;  $\beta$  represents the log-relative rate of mortality associated with a unit increase of air pollutants;  $Z_t$  indicates the pollutant concentrations on day t; DOW is day of the week effect; ns(time,df) is the natural spline function of calendar time; and ns(temperature / humidity,3) is the natural spline function of temperature / humidity with 3 df.

We fitted both single-pollutant models and models with a different combination of pollutants (up to two pollutants per model) to assess the individual effect of each pollutant as well as the stability of the pollutants' effects.

All analyses were conducted in R 2.5.1 using the MGCV package  $^{16)}$ .

#### Results

From 2001 to 2004, a total of 173,911 deaths (91,314 males and 82,597 females) were registered in the study population (Table 1). The percentages of total deaths by age group were 0.3% for 0–4, 3.2% for 5–44, 13.0% for 45–64 and 83.5% for 65+, respectively. On average, there were 119.0 non-accidental deaths per day, including 44.2 from cardiovascular diseases and 14.3 from respiratory

|                  |        | ε .    | ,           |                   |
|------------------|--------|--------|-------------|-------------------|
|                  | $SO_2$ | $NO_2$ | Temperature | Relative humidity |
| $PM_{10}$        | 0.64   | 0.71   | -0.21       | -0.37             |
| $SO_2$<br>$NO_2$ | 1.00   | 0.73   | -0.21       | -0.52             |
| $NO_2$           |        | 1.00   | -0.38       | -0.27             |
| Temperature      |        |        | 1.00        | 0.21              |

**Table 2.** Correlation coefficients between daily air pollutant concentrations and weather conditions in Metropolitan Shanghai (2001–2004)

**Table 3.** Percent increase of mortality outcomes associated with a 10  $\mu$ g/m<sup>3</sup> increase of air pollutant concentrations under single and multiple pollutant models\*

|                  | Model   | Total<br>mortality | Cardiovascular<br>mortality | Respiratory mortality |
|------------------|---|--------------------|-----------------------------|-----------------------|
| PM1 <sub>0</sub> | single-pollutant                                  | 0.26 (0.14, 0.37)  | 0.27 (0.10, 0.44)           | 0.27 (-0.01, 0.56)    |
|                  | adjusted for SO <sub>2</sub>                      | 0.08 (-0.07, 0.22) | 0.12 (-0.10, 0.34)          | -0.04 (-0.41, 0.33)   |
|                  | adjusted for NO <sub>2</sub>                      | 0.01 (-0.14, 0.17) | 0.01 (-0.22, 0.25)          | -0.05 (-0.45, 0.34)   |
|                  | adjusted for SO <sub>2</sub> and NO <sub>2</sub>  | 0.00 (-0.16, 0.16) | 0.01 (-0.23, 0.25)          | -0.10 (-0.50, 0.30)   |
| 2                | single-pollutant                                  | 0.95 (0.62, 1.28)  | 0.91 (0.42, 1.41)           | 1.37 (0.51, 2.23)     |
|                  | adjusted for PM <sub>10</sub>                     | 0.80 (0.37, 1.24)  | 0.69 (0.04, 1.34)           | 1.45 (0.32, 2.59)     |
|                  | adjusted for NO <sub>2</sub>                      | 0.31 (-0.27, 0.89) | 0.05 (-0.83, 0.93)          | 0.95 (-0.59, 2.50)    |
|                  | adjusted for PM <sub>10</sub> and NO <sub>2</sub> | 0.31 (-0.28, 0.89) | 0.04 (-0.85, 0.93)          | 1.00 (-0.55, 2.56)    |
| NO <sub>2</sub>  | single-pollutant                                  | 0.97 (0.66, 1.27)  | 1.01 (0.55, 1.47)           | 1.22 (0.42, 2.01)     |
|                  | adjusted for PM <sub>10</sub>                     | 0.94 (0.50, 1.38)  | 0.98 (0.33, 1.64)           | 1.32 (0.21, 2.45)     |
|                  | adjusted for SO <sub>2</sub>                      | 0.73 (0.19, 1.27)  | 0.97 (0.16, 1.80)           | 0.47 (-0.94, 1.91)    |
|                  | adjusted for PM <sub>10</sub> and SO <sub>2</sub> | 0.73 (0.14, 1.32)  | 0.95 (0.06, 1.84)           | 0.62 (-0.92, 2.16     |

<sup>\*</sup>Current day temperature and relative humidity (lag 0), and two-day moving average of air pollutants concentrations (lag 01) were used in all the regression models of Table 3.

diseases. Cardiopulmonary disease accounted for 49.1% of the total non-accidental deaths for the urban residents of Shanghai.

During our study period, the mean daily average concentrations of  $PM_{10}$ ,  $SO_2$  and  $NO_2$  were  $102.0~\mu g/m^3$ , 44.7  $\mu g/m^3$  and 66.6  $\mu g/m^3$ , respectively. Meanwhile, the mean daily average temperature and relative humidity were 17.7°C and 72.9%, reflecting the subtropical climate in Shanghai.

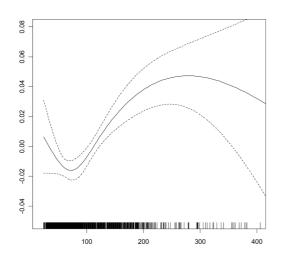
Generally, PM<sub>10</sub>, SO<sub>2</sub> and NO<sub>2</sub> had relatively high correlation coefficients with each other, and PM<sub>10</sub>/SO<sub>2</sub>/NO<sub>2</sub> concentrations were negatively correlated with mean temperature and relative humidity (Table 2).

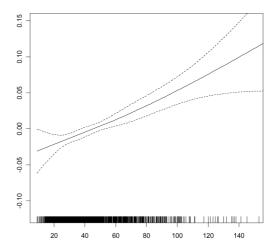
Table 3 shows the comparison results of the single-pollutant models and multiple-pollutant models. In the single-pollutant models, significant association was established between levels of air pollutants (PM<sub>10</sub>, SO<sub>2</sub>, and NO<sub>2</sub>) and daily total non-accidental mortality as well as cardiorespiratory mortality. An increase of 10  $\mu$ g/m³ of PM<sub>10</sub>, SO<sub>2</sub> or NO<sub>2</sub> corresponds to 0.25% (95%CI 0.14–0.37%), 0.95% (95%CI 0.62–1.27%) or 0.96% (95%CI 0.66–1.26%) increases of all-cause mortality, respectively.

Figure 2 graphically shows the exposure-response relationship for each pollutant with total mortality in the single-pollutant models. For most concentration levels (>70  $\mu$ g/m³) of PM<sub>10</sub>, we observed a positive non-linear relationship between PM<sub>10</sub> and three mortality outcomes. There was an almost linear relationship between SO<sub>2</sub> and total mortality. We did not observe any obvious threshold concentration below which SO<sub>2</sub> had no effect on total mortality. There was a positive non-linear relationship between NO<sub>2</sub> and mortality outcomes.

In the multiple-pollutant models, the effect of  $PM_{10}$  on total and cardiovascular morality decreased and became insignificant after adjustment for  $SO_2$ ,  $NO_2$  or both.  $PM_{10}$  has no significant effect on respiratory mortality either before or after adjustment for co-pollutants. The effect of  $SO_2$  on total, cardiovascular and respiratory mortality remained significant after adding  $PM_{10}$  into the models; however, the effect of  $SO_2$  became statistically insignificant for all three mortality outcomes after adjustment for  $NO_2$ , or both  $PM_{10}$  and  $NO_2$ . The effect of  $NO_2$  on total and cardiovascular mortality did not alter much when other pollutants were added. The effect of

a: b:





c:

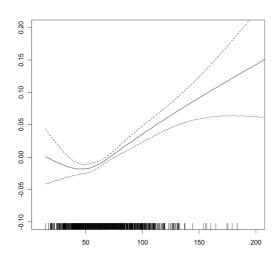


Fig. 2. Smoothing plots of air pollutants against all-cause mortality risk (df=3). X-axis is the 2-day average (lag 01) air pollutants concentrations ( $\mu$ g/m³). The estimated mean percentage of change in daily mortality is shown by the solid line, and the dotted line represents twice the point-wise standard error. (a: PM<sub>10</sub>; b: SO<sub>2</sub>; c: NO<sub>2</sub>)

NO<sub>2</sub> on respiratory mortality was not affected after adjustment for PM<sub>10</sub>, but became insignificant after adjustment for SO<sub>2</sub>, or both PM<sub>10</sub> and SO<sub>2</sub>.

#### Discussion

Evidence gained in this time-series analysis shows that outdoor air pollution ( $PM_{10}$ ,  $SO_2$  and  $NO_2$ ) was associated with mortality from all causes and from cardiopulmonary diseases in Shanghai in 2001–2004. As in prior studies, we found a significant association between  $PM_{10}$  and mortality outcomes in Shanghai, although the size of our estimate was relatively small. We also found a significant effect of gaseous pollutants ( $SO_2$ , and  $NO_2$ ) on daily

mortality, even after adjustment for  $PM_{10}$ . Although  $SO_2$  and  $NO_2$  contribute to PM formation, the current analysis suggests that they are also pollutants that can independently cause adverse health effects.

 $\mathrm{SO}_2$  is a gaseous pollutant produced by fuel combustion, and it can cause bronchoconstriction in normal and asthmatic subjects after short term exposure<sup>17, 18)</sup>.  $\mathrm{SO}_2$  can be converted to sulphuric acid, which can be carried into the small airways by inhalable particulates and impair lung function in children<sup>19)</sup>. Although the high correlation between  $\mathrm{SO}_2$  and co-pollutants such as  $\mathrm{PM}_{10}$  and  $\mathrm{NO}_2$  made it difficult to separate its independent effect, an intervention study conducted in Hong Kong provided direct evidence

that  $SO_2$  resulting from sulfur-rich fuels had an effect on cardio-respiratory deaths<sup>20)</sup>. Also in Hong Kong, the strongest effects of outdoor air pollution were for gases including  $SO_2$  rather than  $PM_{10}^{21}$ . Moreover, it is well known that ambient  $SO_2$  is the precursor of  $SO_4^{2-}$  or sulfate, an important component of fine particles. Zhang et al. found that the concentration of  $SO_4^{2-}$  in the air was closely associated with chronic disease mortality in Beijing, China<sup>22)</sup>.

Of the pollutants we considered, only NO, remained significantly associated with total and cardiovascular mortality after adjustment for co-pollutants. This result is consistent with a recent multi-city analysis in Europe (APHEA-2)<sup>23)</sup>. However, in an analysis of 20 U.S. cities within the National Morbidity, Mortality and Air Pollution Study (NMMAPS), no consistent pattern of association between total mortality and NO<sub>2</sub> was found<sup>6)</sup>. The difference between NMMAPS and APHEA findings may be attributed to the varying air pollution sources and mixture in the US and Europe<sup>23)</sup>. Inhalation of NO<sub>2</sub> may provoke an inflammatory response in the lungs with the consequent release into the circulation of prothrombotic and inflammatory cytokines; a systemic acute phase response of this nature would put people with coronary atheroma at increased risk of plaque rupture and thrombosis. Moreover, exposure to NO, may have an adverse effect on cardiac autonomic control, leading to an increased risk of arrhythmia in susceptible patients<sup>24</sup>). For example, acute exposure to NO<sub>2</sub> has been associated with ventricular tachyarrhythmias which are common precursors to sudden cardiac death<sup>25)</sup>. Of course, NO<sub>2</sub> could be a marker for other pollutants generated from the same source such as particles. For example, Seaton et al suggested NO<sub>2</sub> is a surrogate for the ultra-fine particle (UFP) number<sup>26)</sup>. Therefore, there is a possibility that the effect we observed for NO, might be due to other unmeasured pollutant such as UFP.

Of course, given that the present air pollution epidemiologic study used ambient pollutant concentrations as surrogates of personal exposure, the observed health effects attributed to the ambient gaseous pollutants, e.g. SO<sub>2</sub> and NO<sub>2</sub>, might actually be a result of exposures to fine particles <sup>10, 27, 28</sup>. At present, we cannot conclude that SO<sub>2</sub> and NO<sub>2</sub> are proxies of fine particles or the components of fine particles, and SO<sub>2</sub> and NO<sub>2</sub> may have a direct short-term effect on mortality in Shanghai. However, a consistent, significant effect of SO<sub>2</sub> and NO<sub>2</sub> on mortality observed in China suggests that the role of outdoor exposure to gaseous pollutants should be investigated further.

For air pollution, the shape of the exposure-response relation for mortality has long been of interest as researchers have attempted to supply useful evidence for policymakers seeking to minimize the risk to public health. We did not find a linear exposure-response relationship of total mortality with PM<sub>10</sub> or NO<sub>2</sub> (Fig. 2), which might have contributed to the relatively lower risk of the two pollutants in our study compared with the findings of prior studies. For PM<sub>10</sub>, a "U" relationship with mortality risk was found for most concentrations we observed ( $\leq 300~\mu g/m^3$ ). For NO<sub>2</sub>, the mortality risk seems to increase only with higher concentrations above  $50~\mu g/m^3$ . Some prior studies have suggested the existence of threshold effects in the population risk level due to air pollution exposure<sup>29</sup>). Given the high correlation between pollutant concentrations (Table 2), the linearity of the relationship between air pollution and mortality should be explored further in the future.

In summary, unlike some prior studies in North America, we found significant effects of SO<sub>2</sub> and NO<sub>2</sub> on mortality outcomes even after adjustment for PM<sub>10</sub> in Shanghai. Our findings, combined with previous Chinese studies showing a consistent, significant effect of gaseous pollutants on mortality, suggest that the role of outdoor exposure to SO<sub>2</sub> and NO<sub>2</sub> should be investigated further in China.

Shanghai will host the World Expo in 2010, and plans to invest billions of dollars in environment protection to prepare for this momentous event. The results of the current study may supplement useful information on air pollution-related health effects in Shanghai, thereby providing local decision-makers with information needed to set priority of air pollution control measures with the largest health benefits.

Acknowledgment: The study was funded by the U.S. Health Effects Institute through Grant 4717-RFIQ03-3/04-13. Haidong Kan has been personally supported by Rising Star Program for Young Investigators of Shanghai Municipal Committee of Science and Technology since 2004 (04QMX1402).

The views expressed in this article are those of the authors and do not necessarily reflect the views of the Health Effects Institute or its sponsors.

The authors declare they have no competing financial interests.

## Reference

- 1) Goldberg MS, Burnett RT and Stieb D: A review of time-series studies used to evaluate the short-term effects of air pollution on human health. Rev Environ Health 18, 269–303 (2003)
- Burnett RT, Brook J, Dann T, Delocla C, Philips O, Cakmak S, Vincent R, Goldberg MS and Krewski D: Association between particulate- and gas-phase components of urban air pollution and daily mortality in eight Canadian cities. Inhal Toxicol 12, S4, 15–39 (2000)
- Katsouyanni K, Schwartz J, Spix C, Touloumi G, Zmirou D, Zanobetti A, Wojtyniak B, Vonk JM, Tobias A, Ponka A, Medina S, Bacharova L and Anderson

- HR: Short term effects of air pollution on health: a European approach using epidemiologic time series data: the APHEA protocol. J Epidemiol Community Health 50, S1, 12–18 (1996)
- 4) Katsouyanni K, Touloumi G, Samoli E, Gryparis A, Le Tertre A, Monopolis Y, Rossi G, Zmirou D, Ballester F, Boumghar A, Anderson HR, Wojtyniak B, Paldy A, Braunstein R, Pekkanen J, Schindler C and Schwartz J: Confounding and effect modification in the shortterm effects of ambient particles on total mortality: results from 29 European cities within the APHEA2 project. Epidemiology 12, 521–531 (2001)
- 5) Katsouyanni K, Touloumi G, Spix C, Schwartz J, Balducci F, Medina S, Rossi G, Wojtyniak B, Sunyer J, Bacharova L, Schouten JP, Ponka A and Anderson HR: Short-term effects of ambient sulphur dioxide and particulate matter on mortality in 12 European cities: results from time series data from the APHEA project. Air Pollution and Health: a European Approach. BMJ 314, 1658–1663 (1997)
- Samet JM, Dominici F, Curriero FC, Coursac I and Zeger SL: Fine particulate air pollution and mortality in 20 U.S. cities, 1987–1994. N Engl J Med 343, 1742– 1749 (2000)
- Health Effects Institute. In: Health effects of outdoor air pollution in developing countries of Asia: a literature review. Boston: Health Effects Institute, 2004.
- Derriennic F, Richardson S, Mollie A and Lellouch J: Short-term effects of sulphur dioxide pollution on mortality in two French cities. Int J Epidemiol 18, 186– 197 (1989)
- Xu X, Gao J, Dockery DW and Chen Y: Air pollution and daily mortality in residential areas of Beijing, China. Arch Environ Health 49, 216–222 (1994)
- 10) Sarnat JA, Schwartz J, Catalano PJ and Suh HH: Gaseous pollutants in particulate matter epidemiology: confounders or surrogates? Environ Health Perspect 109, 1053–1061 (2001)
- Zeger SL, Irizarry R and Peng RD: On time series analysis of public health and biomedical data. Annu Rev Public Health, 27, 57–79 (2006)
- Hastie TJ and Tibshirani RJ, eds. Generalized Additive Models. London: Chapman & Hall, 1990.
- Peng RD, Dominici F and Louis TA: Model choice in time series studies of air pollution and mortality. J R Stat Soc A Stat 169, 179–203 (2006)
- 14) Bell ML, McDermott A, Zeger SL, Samet JM and Dominici F: Ozone and short-term mortality in 95 US urban communities, 1987–2000. JAMA 292, 2372–2378 (2004)
- 15) Dominici F, Peng RD, Bell ML, Pham L, McDermott A, Zeger SL and Samet JM: Fine particulate air pollution and hospital admission for cardiovascular and respiratory diseases. JAMA 295, 1127–1134 (2006)
- 16) R Development Core Team: R: A language and environment for statistical computing. Vienna: R

- Foundation for Statistical Computing, 2005.
- 17) Sheppard D, Wong WS, Uehara CF, Nadel JA and Boushey HA: Lower threshold and greater bronchomotor responsiveness of asthmatic subjects to sulfur dioxide. Am Rev Respir Dis 122, 873-878 (1980)
- 18) Stacy RW, Seal E Jr., House DE, Green J, Roger LJ and Raggio L: A survey of effects of gaseous and aerosol pollutants on pulmonary function of normal males. Arch Environ Health 38, 104–115 (1983)
- 19) Spengler JD, Koutrakis P, Dockery DW, Raizenne M and Speizer FE: Health effects of acid aerosols on North American children: air pollution exposures. Environ Health Perspect 104, 492–499 (1996)
- 20) Hedley AJ, Wong CM, Thach TQ, Ma S, Lam TH and Anderson HR: Cardiorespiratory and all-cause mortality after restrictions on sulphur content of fuel in Hong Kong: an intervention study. Lancet 360, 1646–1652 (2002)
- 21) Wong CM, Ma S, Hedley AJ and Lam TH: Effect of air pollution on daily mortality in Hong Kong. Environ Health Perspect 109, 335–340 (2001)
- 22) Zhang J, Song H, Tong S, Li L, Liu B and Wan L: Ambient sulfate concentration and chronic disease mortality in Beijing. Sci Total Environ 262, 63–71 (2000)
- 23) Samoli E, Aga E, Touloumi G, Nisiotis K, Forsberg B, Lefranc A, Pekkanen J, Wojtyniak B, Schindler C and Niciu E: Short-term effects of nitrogen dioxide on mortality: an analysis within the APHEA project. Eur Respir J 27, 1129–1138 (2006)
- 24) Routledge HC, Ayres JG and Townend JN: Why cardiologists should be interested in air pollution. Heart 89, 1383–1388 (2003)
- 25) Dockery DW, Luttmann-Gibson H, Rich DQ, Link MS, Mittleman MA, Gold DR, Koutrakis P, Schwartz JD and Verrier RL: Association of air pollution with increased incidence of ventricular tachyarrhythmias recorded by implanted cardioverter defibrillators. Environ Health Perspect 113, 670–674 (2005)
- 26) Seaton A and Dennekamp M: Hypothesis: ill health associated with low concentrations of nitrogen dioxide—an effect of ultrafine particles? Thorax 58, 1012–1015 (2003)
- 27) Sarnat JA, Brown KW, Schwartz J, Coull BA and Koutrakis P: Ambient gas concentrations and personal particulate matter exposures: implications for studying the health effects of particles. Epidemiology 16, 385–395 (2005)
- Schwartz J: Total suspended particulate matter and daily mortality in Cincinnati, Ohio. Environ Health Perspect 102, 186–189 (1994)
- 29) Kim SY, Lee JT, Hong YC, Ahn KJ and Kim H: Determining the threshold effect of ozone on daily mortality: an analysis of ozone and mortality in Seoul, Korea, 1995–1999. Environ Res 94, 113–119 (2004)